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Original Manuscript

Interspecific differences in oxidative DNA damage after hydrogen peroxide exposure of sea urchin coelomocytes

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Abstract

Interspecific comparison of DNA damage can provide information on the relative vulnerability of marine organisms to toxicants that induce oxidative genotoxicity. Hydrogen peroxide (H₂O₂) is an oxidative toxicant that causes DNA strand breaks and nucleotide oxidation and is used in multiple industries including Atlantic salmon aquaculture to treat infestations of ectoparasitic sea lice. H₂O₂ (up to 100 mM) can be released into the water after sea lice treatment, with potential consequences of exposure in nontarget marine organisms. The objective of the current study was to measure and compare differences in levels of H₂O₂-induced oxidative DNA damage in coelomocytes from Scottish sea urchins *Echinus esculentus*, *Paracentrotus lividus*, and *Psammechinus miliaris*. Coelomocytes were exposed to H₂O₂ (0–50 mM) for 10 min, cell concentration and viability were quantified, and DNA damage was measured by the fast micromethod, an alkaline unwinding DNA method, and the modified fast micromethod with nucleotide-specific enzymes. Cell viability was >92% in all exposures and did not differ from controls. *Psammechinus miliaris* coelomocytes had the highest oxidative DNA damage with 0.07 ± 0.01, 0.08 ± 0.01, and 0.07 ± 0.01 strand scission factors (mean ± SD) after incubation with phosphate-buffered saline, formamidopyrimidine-DNA glycosylase, and endonuclease-III, respectively, at 50 mM H₂O₂. Exposures to 0.5 mM H₂O₂ (100-fold dilution from recommended lice treatment concentration) induced oxidative DNA damage in all three species of sea urchins, suggesting interspecific differences in vulnerabilities to DNA damage and/or DNA repair mechanisms. Understanding impacts of environmental genotoxicants requires understanding species-specific susceptibilities to DNA damage, which can impact long-term stability in sea urchin populations in proximity to aquaculture farms.

Keywords: DNA strand breaks; DNA damage; environmental effects; environmental toxicology; genotoxic effects in different organisms; genotoxicity

Introduction

The use of hydrogen peroxide (H₂O₂) as an antiparasitic treatment in Atlantic salmon aquaculture leads to dispersion into the marine environment and unintended exposure of surrounding marine organisms. The recommended treatment concentration and duration are 1.2–1.8 g/l (35–53 mM) H₂O₂ for 20 min [1], which induces a mechanical paralysis and detachment of the sea louse from the fish [2]. H₂O₂ is estimated to degrade into oxygen and water rapidly in just over 7 days [3], and is therefore considered to be the most environmentally friendly chemical option for sea lice treatment [4]. Although H₂O₂ is expected to dilute after release, hydrodynamic modelling has found that concentrations of 100–1000 times dilutions (530–53 μM) can persist over one kilometre from the release site, with speed and extent of dilution dependent on local weather and current conditions [5, 6]. H₂O₂ is denser than seawater and will sink to the seabed [7, 8]. The plume behaviour of H₂O₂ after a treatment event is therefore complex and dynamic, and both pelagic and benthic organisms are vulnerable to be exposed to elevated levels of H₂O₂ for hours or even days [8].

DNA damage can occur by exposure to environmental toxicants and lead to genome instability. Oxidative substances such as H₂O₂ that produce reactive oxygen species can cause oxidative DNA damage, including double- and single-stranded DNA breaks [4, 9–14]. Double- and single-stranded DNA breaks are the most harmful forms of DNA damage [13]. Unrepaired or misrepaired DNA strand breaks can lead to cell death, chromosome translocation, and genomic instability [15]. DNA strand breaks are often caused by oxygen radicals by inhibition in the progression of RNA polymerase [16]. Single-strand breaks can also deteriorate into the more lethal double-stranded DNA breaks by replication forks collapsing at sites of the break during replication [17]. In addition to DNA strand breaks, H₂O₂ can directly oxidize purine and pyrimidine deoxyribonucleotides where the former are more susceptible to oxidation than the latter as guanine has the lowest oxidation potential of all four nucleobases [18]. Unrepaired oxidized purine and pyrimidine lesions can lead to mutagenesis, cytostasis, and cytotoxicity [19].

Marine organisms including sea urchins share the coastal environment with anthropogenic activities; therefore, they can

be useful models for assessment of genotoxic impacts from discharge of contaminants into the surrounding ecosystem. Sea urchins are found all over the world with extensive near-shore salmon aquaculture in Scotland, Chile, and Norway. Several species of sea urchins are found in abundance in proximity of salmon farms on the west coast of Scotland, with access to salmon cage waste [20, 21]. With the release of H_2O_2 from salmon farms, slower moving sea urchins can be directly impacted by waste plumes, making them a useful bioindicator for sea lice treatment effects. Those inhabiting coastal areas can have ecological roles influenced by human activities and previous studies have used them as bioindicators for local pollutants [22–24]. They are pivotal organisms in the maintenance of ecological balance in the marine ecosystem and are found in all marine environments [25, 26]. European lobsters at pelagic larvae stage [27], deep-water shrimp [8], zooplankton [28], and krill [29] have all been observed to suffer adverse behavioural and genotoxic effects, when exposed to 10–1000 times dilutions (53–5.3 mM) of the recommended sea lice treatment, and sublethal behavioural changes observed at 100–200 times dilutions (265–530 μ M). It is unknown how sea urchin coelomocytes will respond to realistic environmental concentrations of H_2O_2 , however their role as a benthic bioindicator species [30], along with their ease of collection, maintenance, husbandry, cell extraction, transparency and large quantities of embryos, and larvae, make them an ideal model for investigating H_2O_2 impacts.

There are several methods for DNA damage detection and measurements. The fast micromethod detects single- and double-stranded DNA breaks, and alkali labile sites by means of alkaline unwinding, following the same first principles of alkaline DNA unwinding in the comet assay, but with more scope for other cell systems, sensitivity, larger sample sets, and more rapid quantification [13, 31, 32]. Studies have used the fast micromethod to detect DNA damage in sea urchin coelomocytes after short exposures to H_2O_2 [12, 33, 34], and the current study looks to further investigate oxidized purine and pyrimidine nucleotides caused by H_2O_2 in addition to strand breaks. Previously, the comet assay was modified to incorporate enzymes into the protocol for lesion-specific DNA damage detection [35]. Formamidopyrimidine-DNA glycosylase (FPG) and endonuclease-III (Endo III) are bacterial repair enzymes used in the modified comet assay for the measurement of oxidized purines and pyrimidines and they have been identified as efficient biomarkers for detecting oxidative DNA damage [36, 37]. FPG and Endo III work by recognizing and cleaving oxidized purine and pyrimidine nucleotides, respectively, converting the oxidative damage into single strand breaks [38]. Compared with the comet assay, the fast micromethod is a more rapid method for determining DNA damage suitable for large quantity of samples in ecotoxicological studies [31]. Whereas the comet assay separates and spreads unwound DNA by electrophoresis, the fast micromethod incorporates PicoGreen which preferentially binds to double-stranded DNA and is kinetically released during the alkaline process of unwinding, allowing for quantification of the rate of release (reduction in fluorescence) correlated with initial levels of DNA strand breaks.

The aim of the current study was to measure and compare levels of oxidative DNA damage in coelomocytes of cold-water sea urchin species *Echinus esculentus*, *Paracentrotus lividus*, and *Psammechinus miliaris* from the west coast of

Scotland and within the vicinity of Atlantic salmon aquaculture. The objectives were: (i) develop enzyme modified fast micromethod for investigating lesion-specific DNA damage in coelomocytes; (ii) identify and differentiate oxidative DNA damage susceptibility between the three sea urchins species; (iii) differentiate and quantify H_2O_2 -induced DNA strand breaks, oxidized purines, and oxidized pyrimidines; and (iv) interpret results in the context of interspecific differences in size, coelomocyte characteristics, and within the context of environmental exposure to H_2O_2 antiparasitic treatments in coastal Atlantic salmon aquaculture.

Materials and methods

Animal maintenance

Adult stocks of *E. esculentus*, *P. lividus*, and *P. miliaris* were maintained at the Scottish Association for Marine Science (SAMS) aquarium in a flow-through system with filtered natural seawater, ambient temperature, and photoperiod, and were fed *ad libitum* with macroalgae. Sea urchins used in the current study were selected from adult stocks kept in the aquarium for up to 9 years. Animals were maintained and handled in accordance with UK animal welfare regulations. Routine aquarium husbandry and experimental protocols received ethical approval from University of Highlands and Islands research ethics committee and animal welfare and ethics committee (application ID: 503). Each sea urchin was measured to determine test diameter and height before coelomic fluid extraction, and in total, 45 of *E. esculentus*, 31 of *P. lividus*, and 12 of *P. miliaris* were used in the current study.

Coelomocyte extraction and H_2O_2 exposure

Coelomocytes were collected by sublethal extraction of coelomic fluid (0.5–1 ml) with a syringe and an 18- or 21-gauge needle inserted at a 45° angle through the peristomial membrane surrounding the Aristotle's lantern. Syringes were pre-filled with ice-cold calcium/magnesium-free seawater (460 mM NaCl, 10 mM KCl, 7 mM Na_2SO_4 , 2.4 mM $NaHCO_3$) containing 30 mM ethylenediaminetetraacetic acid (EDTA) anticoagulant [pH 7.4, 1:1 (vol:vol), with coelomic fluid] to prevent cells clumping [39]. Cell concentration and differential cell counts (red spherule or clear cells) were determined with a Neubauer haemocytometer. Cells were diluted to 2.78 M/ml with 30 mM EDTA anticoagulant. In total, coelomocyte concentrations from $n = 45$ *E. esculentus*, $n = 31$ *P. lividus*, and $n = 12$ *P. miliaris* were recorded, coelomocytes from $n = 8$ *E. esculentus*, $n = 4$ *P. lividus*, and $n = 4$ *P. miliaris* were used for the fast micromethod, and coelomocytes from $n = 6$ *E. esculentus*, $n = 6$ *P. lividus*, and $n = 6$ *P. miliaris* were used for the modified fast micromethod.

Coelomocytes were exposed to H_2O_2 [35% stabilized H_2O_2 (w/w), VWR international Ltd, Leicestershire, UK] at 1:10 ratio to reach final H_2O_2 concentrations of 0.5–50 mM, PBS at 1:10 ratio was added to control coelomocytes. Samples were exposed for 10 min on ice and in the dark, same as acute H_2O_2 exposures in previous research on DNA damage in sea urchins [33]. After exposure, subsamples were placed on a Neubauer haemocytometer with 0.5% trypan blue at 1:1 ratio, and cell viabilities were assessed by counting of stained blue dead cells and clear live cells.

DNA damage detection

After exposure to H₂O₂, coelomocytes were aliquoted into quadruplicate wells (96-well black-walled microplates, Greiner Bio-One Ltd) with cell concentration of ~50 000 cells per well, consistent with previously published method from Reinardy and Bodnar [12]. Lysis solution (9 M urea, 0.1% SDS, 0.2 M EDTA, pH 10) with PicoGreen (1:50, Fisher Scientific, Leicestershire, UK) was added to each well, plate planks contained PBS without coelomocytes, and lysed in the dark on ice for 40 min. To initiate DNA unwinding, 200 µl unwinding solution (1 M NaOH, 20 mM EDTA, pH 13) was added to each well, fluorescence was immediately read by kinetic mode every 5 min for 20 min (POLARstar Omega, excitation wavelength at 480 nm and emission wavelength at 520 nm, BMG LABTECH Ltd, Bucks, UK).

The fast micromethod was modified to include enzyme-specific incubations for differentiation of oxidized purines and pyrimidines, following the principle established for the modified comet assay [36, 40]. Following separate H₂O₂ exposure experiments, lysis solution was added to the coelomocytes in 1.5 ml microcentrifuge tubes at 1:1 (vol:vol) ratio and cells were lysed in the dark and on ice for 40 min. After lysis, each sample was split into three subsamples and were either incubated with PBS (no enzyme set), FPG (0.04 units), or Endo III (0.05 units) (New England Biolabs, Herts, UK). FPG and Endo III were diluted at 1:1000 with Milli-Q water. Enzymes or PBS control were added to the coelomocytes, followed by incubation in a thermo cycler (MJ Research, Inc., PTC-100TM Programmable Thermo Controller) for 30 min at 37°C and 40 min deactivation step at 4°C [40, 41]. After incubation, cells (approx. 50 000 cells/well) were loaded into quadruplicate wells. PicoGreen, diluted with PBS with ratio of 1:10, was added to each reaction for the fluorescence binding of double-stranded DNA and allowed to sit on ice for 10 min. DNA unwinding was initiated and fluorescence was read same as the fast micromethod.

Data analysis

Total cell concentration, red cell percentage, cell viability, test diameter, and height datasets in *E. esculentus*, *P. lividus*, and *P. miliaris* were tested with one variable analysis Shapiro–Wilk test for normal distribution. Total cell concentrations and test diameters for all species were normally distributed. Red cell percentages were normally distributed for *P. miliaris* but not normally distributed for *E. esculentus* and *P. lividus*. Test heights were normally distributed for *P. lividus* and

P. miliaris, but not normally distributed for *E. esculentus*. Cell viabilities were not normally distributed for all species. Nonparametric Kruskal–Wallis one-way analysis of variance on ranks (ANOVA on Ranks) followed by a *post hoc* Dunn’s method was used to compare differences between species for datasets that were not normally distributed, and one-way ANOVA was used to compare differences between species for datasets that were normally distributed.

DNA damage measured by fast micromethod was calculated by the strand scission factor (SSF) equation: $SSF = \log(\% \text{ dsDNA}_{\text{sample}} / \% \text{ dsDNA}_{\text{control}}) \times (-1)$ [12, 31, 33]. Fast micromethod data were normally distributed and tested with ANOVA for concentration-dependent increases of DNA damage (SSF). Concentration-dependent DNA damages were also modelled by three-parameter logistic regression $f = a / (1 + \exp(-(x - x_0)^b))$ for the fast micromethod [12]. Modified fast micromethod data were normally distributed and statistically tested with general linear model (GLM) for concentration-dependent increases of SSF, and ANOVA to determine the differences between no enzyme, FPG, and Endo III sets of data at different concentrations of H₂O₂. Three-parameter logistic regression was also performed on no enzyme, FPG, and Endo III datasets separately for statistical testing of concentration-dependent DNA damage responses. All statistical analyses were performed in SigmaPlot 14.0 (Systat Software, Inc.) or Statgraphics 19 (Statgraphics Technologies, Inc., VA, USA).

Results

Out of the three species of sea urchins, *P. miliaris* was the smallest ($P < .05$, ANOVA or Kruskal–Wallis) with diameter of 39 ± 3.5 (means \pm SD, $n = 12$) mm and height of 21 ± 3.0 (means \pm SD, $n = 12$) mm (Table 1). No significant differences in size were found between *E. esculentus*, diameter of 46 ± 4.6 mm and height of 24 ± 4.2 mm, and *P. lividus*, diameter of 46 ± 9.4 mm and height of 24 ± 5.1 mm ($P > .05$, ANOVA or Kruskal–Wallis). All three species had total coelomocytes concentrations of $10.24\text{--}13.02 \times 10^6$ cells/ml and red spherule cells percentage of 4.2%–9.5% (means \pm SD) (Table 1). No significant differences in total cell concentrations were found ($P > .05$, ANOVA), but *E. esculentus* had a significantly lower percentage of red cells of $4.2 \pm 4.3\%$ compared with *P. lividus* and *P. miliaris* ($P < .05$, ANOVA). Red cell percentages of *P. lividus*, $6.9 \pm 3.8\%$, and *P. miliaris*, $9.5 \pm 4.9\%$, were not significantly different ($P > .05$, Kruskal–Wallis). Cell viabilities, as recorded in fast micromethod and modified fast micromethod in coelomocytes in all species

Table 1. Sea urchin sizes and coelomocytes concentrations.

Sea urchin species	<i>n</i>	Test diameter (mm)	Test height (mm)	Red spherule cells (% of total coelomocyte concentration)	Total coelomocyte concentration ($\times 10^6$ cells/ml)	Cell viability at the highest H ₂ O ₂ concentration (%)
<i>E. esculentus</i>	45	46 \pm 4.6	24 \pm 4.2	4.2 \pm 4.3*	11.8 \pm 4.2	96.6 \pm 2.6
<i>P. lividus</i>	31	46 \pm 9.4	24 \pm 5.1	6.9 \pm 3.8	13.02 \pm 5.8	97.3 \pm 1.0
<i>P. miliaris</i>	12	39 \pm 3.5*	21 \pm 3.0*	9.5 \pm 4.9	10.24 \pm 2.7	98.2 \pm 1.0

E. esculentus ($n = 45$), *P. lividus* ($n = 31$), and *P. miliaris* ($n = 12$) test diameter and height shown in mean \pm SD mm. Total cell concentration shown in mean \pm SD $\times 10^6$ cells/ml and red spherule cells in mean \pm SD % of total coelomocyte concentration. Cell viability assessed with 0.5% trypan blue on coelomocytes exposed to 50 mM H₂O₂ and shown in mean \pm SD %.

*Significant differences in test diameter and height, and red spherule cells % ($P < .05$, ANOVA or Kruskal–Wallis).

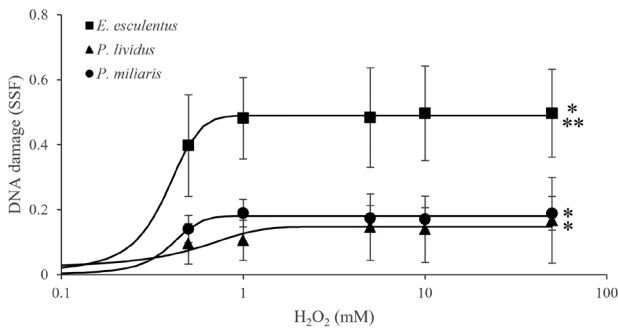


Figure 1. H_2O_2 concentration-dependent DNA damage (SSF, fast micromethod) of coelomocytes in three species of sea urchins, *E. esculentus* (■, $n = 8$), *P. lividus* (▲, $n = 4$), and *P. miliaris* (●, $n = 4$). Data were modelled with three-parameter logistic regression and data points are means \pm SD ($n = 4$ –8). *Significant concentration-dependent increase in DNA damage ($P < .05$, ANOVA). **Significant species difference in overall level of SSF ($P < .05$, ANOVA).

exposed to 50 mM H_2O_2 was over 92% (Table 1) with no significant difference in cell viabilities between species ($P > .05$, Kruskal–Wallis).

Echinus esculentus ($n = 8$), *P. lividus* ($n = 4$), and *P. miliaris* ($n = 4$) coelomocytes DNA damage (SSF) increased after exposure to 0–50 mM H_2O_2 (Fig. 1). At 50 mM of H_2O_2 , *P. lividus* and *P. miliaris* had a maximum mean level of DNA damage of 0.17 ± 0.13 and 0.19 ± 0.05 SSFs, respectively, whereas *E. esculentus* had the highest DNA damage at 0.50 ± 0.14 SSF (Fig. 1). *Echinus esculentus* also had significantly the highest SSFs of all three species with 0.40 ± 0.16 , 0.48 ± 0.13 , 0.48 ± 0.15 , 0.50 ± 0.14 , and 0.50 ± 0.14 SSFs after exposure to 0.5, 1, 5, 10, and 50 mM of H_2O_2 ($P < .05$, ANOVA), respectively, while no significant difference in SSFs was found between *P. lividus* and *P. miliaris* ($P > .05$, ANOVA). Acute exposures to 0.5–50 mM H_2O_2 significantly induced DNA damage in *E. esculentus* and *P. miliaris* ($P < .05$, ANOVA), and 5–50 mM of H_2O_2 significantly induced DNA damage in *P. lividus* ($P < .05$, ANOVA).

Coelomocytes exposed to H_2O_2 (0–50 mM) from *E. esculentus* ($n = 6$), *P. lividus* ($n = 6$), and *P. miliaris* ($n = 6$) then incubated with no enzyme, FPG, and Endo III showed significant H_2O_2 -induced oxidative DNA damage ($P < .05$, GLM, Fig. 2). Concentration-dependent DNA damage responses fitted three-parameter logistic regression models significantly ($P < .05$, logistic regression) for no enzyme, FPG, and Endo III incubated coelomocytes from all species of sea urchin. Significant differences in enzyme treatments were found in *P. lividus* at 10 mM H_2O_2 between FPG and Endo III ($P < .05$, GLM), *P. miliaris* at 5 mM H_2O_2 between no enzyme, FPG, and Endo III ($P < .05$, GLM), and *P. miliaris* at 10 mM H_2O_2 between FPG and Endo III ($P < .05$, GLM). In *E. esculentus* at 5 mM H_2O_2 , coelomocytes with FPG and Endo III show higher DNA damage compared with coelomocytes with no enzymes. At 10 mM H_2O_2 , coelomocytes with no enzyme and FPG show similar SSF, whereas coelomocytes with Endo III had lower SSF. At 50 mM H_2O_2 , all three sample sets showed similar SSF. DNA damage reached a plateau at 10 and 50 mM H_2O_2 (Fig. 2A). No significant DNA damage was observed between coelomocytes incubated with no enzyme, FPG, and Endo III for all concentrations of H_2O_2 in *E. esculentus* ($P > .05$, GLM, Fig. 2A). *Paracentrotus lividus* coelomocytes incubated with no enzyme showed lower but nonsignificant

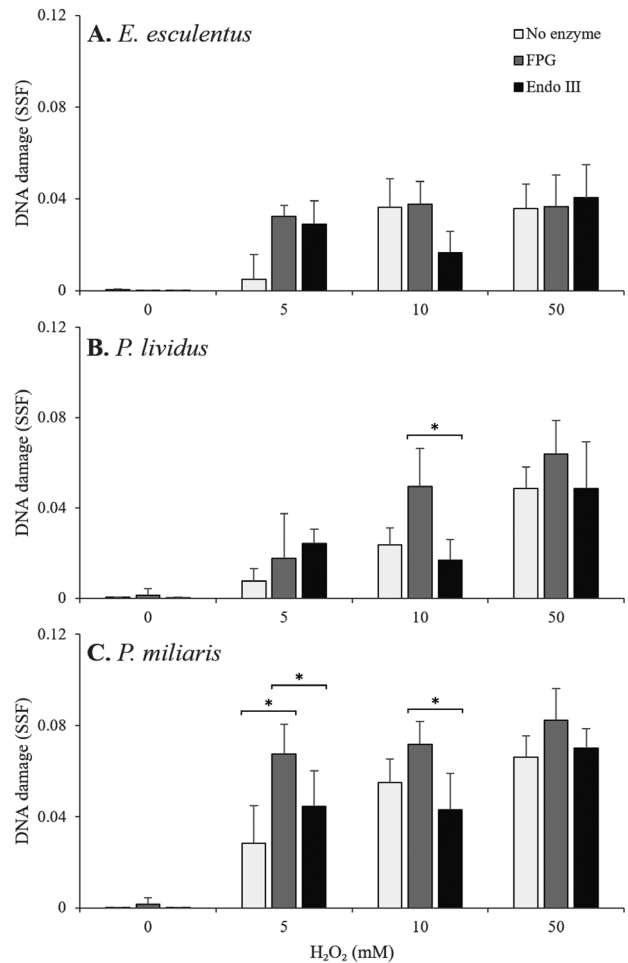


Figure 2. DNA damage (SSF) in *E. esculentus* ($n = 6$, A), *P. lividus* ($n = 6$, B), and *P. miliaris* ($n = 6$, C) coelomocytes exposed to H_2O_2 . Fast micromethod modified to include either no enzyme (white bars), incubation with FPG (grey bars), or incubated with Endo III (black bars). Data are means \pm SD. *Significant difference in SSF between no enzyme, FPG, and Endo III ($P < .05$, GLM). DNA damage SSF was significantly affected by H_2O_2 concentrations for all treatments ($P < .05$, GLM).

SSF in 5 mM H_2O_2 compared with samples with FPG and Endo III at ($P > .05$, GLM, Fig. 2B). Coelomocytes incubated with Endo III had significantly lower SSFs ($P < .05$, GLM) compared with coelomocytes incubated with no enzymes and FPG when treated with 10 mM H_2O_2 (Fig. 2B). At 50 mM, coelomocytes incubated with FPG showed the highest SSF compared with no enzyme and FPG, but it was not significant ($P > .05$, GLM, Fig. 2B). *Paracentrotus lividus* coelomocytes for all enzyme treatments also showed highest SSFs at 50 mM, but not significantly different than at 5 or 10 mM ($P > .05$, GLM, Fig. 2B). *Psammechinus miliaris* showed the highest overall damage levels of SSFs of the three species of sea urchins with concentration-dependent DNA damage (Fig. 2C). Coelomocytes incubated with no enzyme had significantly lower SSF ($P < .05$, GLM) at 5 mM H_2O_2 and coelomocytes incubated with Endo III had significantly lower SSF ($P < .05$, GLM) at 10 mM H_2O_2 (Fig. 2C). *Psammechinus miliaris* coelomocytes incubated with FPG had the highest DNA damage at every treatment concentration of H_2O_2 , similar to *E. esculentus* at 5 and 10 mM H_2O_2 and *P. lividus* at 10

and 50 mM H₂O₂. At both 5 and 10 mM of H₂O₂, significant differences in DNA damage were observed between FPG and Endo III, and between no enzyme and FPG at 5 mM H₂O₂ ($P < .05$, GLM, Fig. 2C).

Discussion

Total cell concentration and percentage of red cells recorded in the current study indicated large variations between individual sea urchins and different species having similar ranges in concentrations. Sea urchins have a complex innate immune system where coelomocytes play a key immunological role and the current study differentiated coelomocytes into red spherule cells and clear cells (white spherule cell, phagocytes, and vibratile cells) [39, 42, 43]. Cell concentrations reported in the current study were higher than previous studies. Both *P. lividus* and *P. miliaris* reported to have 6–8 × 10⁶ cells/ml, whereas the current study reports the two species with averages of 13.02 × 10⁶ cells/ml and 10.24 × 10⁶ cells/ml, respectively [23, 44, 45]. In addition, previous research suggests that sex and proximity to spawning season contributes to the amount of cells present as female and prespawning sea urchins had higher coelomocyte concentrations at 8.2 ± 1.2 × 10⁶ cells/ml and 7.3 ± 1.7 × 10⁶ cells/ml for females and males prespawning, compared with 7.1 ± 2.1 × 10⁵ cells/ml and 6.8 ± 2.7 × 10⁵ cells/ml for females and males postspawning [44]. The current study extracted the coelomocytes from prespawning sea urchins in the winter months and did not distinguish the sexes of sea urchins, which could be contributing to the large variations in coelomocytes concentrations. *Psammechinus miliaris* had the highest average percentage (9.5%) of red cells, which is consistent with previous studies [45], compared with *E. esculentus* (4.2%) and *P. lividus* (6.9%). Coelomocytes function similar to vertebrate blood cells as the main line of defence against wounding and infections by opsonization, coagulation, encapsulation, phagocytosis, clearing foreign particles, and oxygen transport [39, 46–48]. Red spherule cells contain echinochrome A, which has antibacterial and antifungal activities and have been used as a biomarker for heat stress [49, 50]. Sea urchin coelomocytes are known to vary the proportions of red to white coelomocytes between species and individuals according to size and physiological conditions, and the number of red spherule cells can increase when exposed to polluted seawater or injuries [51]. The three species used in the current study had different average sizes with *P. miliaris* the smallest, which is consistent with previous records [21, 52, 53]. Susceptibility to environmental stress and their natural small sizes or temperature tolerance could also contribute to the relatively high red cell percentage in *P. miliaris*. *Psammechinus miliaris* are distributed across the North Sea, far into Scandinavia, and are tolerant to low temperatures down to 0°C [21]. The water temperature during the time of sampling for the current study was >10°C, possibly warmer than what *P. miliaris* is most comfortable in. *Echinus esculentus* has the lowest red cell percentage which may indicate a reduced capacity to respond to stress, or an overall unstressed state.

For accurate estimation of DNA damage in coelomocytes, cell viability was first assessed to confirm the presence of live coelomocytes. Sea urchin coelomocytes are suggested to be relatively tolerant to high concentrations of genotoxicants and have similar DNA repair mechanisms as vertebrates with a range of repair pathways, suggesting a robust DNA damage

response system [12, 54–56]. Coelomocytes in the current study underwent sublethal H₂O₂ exposures of 0.5–50 mM, equivalent to 1- to 100-fold dilution from recommended sea lice treatment regime within an Atlantic salmon farm [1]. The cell viabilities of coelomocytes exposed to H₂O₂ concentrations of 0 and 50 mM were assessed after every exposure and coelomocytes from every sea urchin had cell viability of >92%. No significant differences were observed in cell viabilities between species of sea urchins. Coelomocytes acutely exposed to 10 mM H₂O₂ had an average of 99% viability and modelled LC₅₀ at 120.1 mM, compared with 21% average viability and modelled LC₅₀ at 0.6 mM for sea urchin larvae [12]. Human cell systems have shown 5%–60% cell viability at 10 mM and <5% at 400 μM H₂O₂ exposure for 24 h [57, 58]. *Poeciliopsis* fish hepatoma cells showed <20%, <20%, and <10% cell viabilities after exposure to 0.1, 1, and 10 mM of H₂O₂ for 24 h [59]. Little is known on the range of acute responses to H₂O₂ across marine invertebrate groups and compared with sea urchin cell viabilities. With the minimum of 92% cell viability at a high H₂O₂ concentration of 50 mM, sea urchin coelomocytes are relatively tolerant to genotoxicity, and the results confirm that exposures were sublethal and DNA damage assessments were conducted in live coelomocytes.

The fast micromethod detected significant DNA damage in sea urchin coelomocytes exposed to 0.5–50 mM H₂O₂. DNA damage reached the maximum levels at 1 mM H₂O₂ for all species, beyond which, there was no increase in damage. Previous research showed highest DNA damage from H₂O₂ at ≥1 mM in both adult coelomocytes and larvae reaching maximum SSFs at 1 mM H₂O₂ in the subtropical sea urchin *Lytechinus variegatus* [12]. Reinardy and Bodnar [12] followed the fast micromethod SSF, providing comparable data to the present study, and reported SSFs of the same range and magnitude, and maximum 0.5 SSF, similar range was found in *E. esculentus* from the current study, but higher than what was detected in *P. lividus* and *P. miliaris*. This may be explained by interspecific differences, geographical ranges and environments, methodological differences used i.e. different fluorescence spectrometers, or differences in the pH of alkaline unwinding solutions with pH 12.4 for *L. variegatus* DNA unwinding and pH 13 in the present study. A study using the fast micromethod in coelomocytes from multiple species of echinoderms reported levels of DNA damage ranging from 0.5 to 1.5 SSF across three difference species, and suggested longer-lived echinoderms may invest more in DNA damage tolerance compared with short-lived species [33]. Both Reinardy and Bodnar [12] and El-Bibany *et al.* [33] investigated sea urchin species from subtropical environment in Bermuda as opposed to the cold-water species from the current study, and they quantified levels of DNA damage up to 1.5 SSF, suggesting species from different geographical and thermal conditions may have differences in DNA damage susceptibility and capacity to respond to environmental genotoxicants.

H₂O₂-induced oxidative DNA damage was detected with the modified fast micromethod with FPG and Endo III incubated coelomocytes, indicating H₂O₂ caused DNA strand breaks in addition to direct oxidation of purines and pyrimidines. The modified fast micromethod was developed to detect oxidized DNA damage caused by H₂O₂ with added FPG and Endo III enzyme incubations. It was able to detect, quantify, and differentiate oxidized purine and pyrimidine nucleotides

as well as DNA strand breaks in coelomocytes from different species of sea urchins and sensitive enough to differentiate between species and H_2O_2 concentrations. Compared with the modified comet assay, the modified fast micromethod is able to be performed more rapidly and work on multiple tissue and cell types, and large sample sizes, with the further potential to apply to DNA extracted from any species or tissue type. FPG incubated coelomocytes resulted in the highest SSF across all species, suggesting that H_2O_2 caused more oxidized purine nucleotides than oxidized pyrimidine nucleotides. Previous studies that used FPG and Endo III modified comet assay indicated and support purines being more susceptible to oxidation than pyrimidines [60, 61]. Coelomocytes where no significance was found between no enzyme and FPG incubations suggest the DNA damage caused by H_2O_2 mainly consisted of single-, double-stranded DNA breaks, and alkali labile sites. The current study provides strong evidence that H_2O_2 mainly causes strand breaks and more oxidized purines than pyrimidines. *Echinus esculentus* had the highest DNA damage detected by the fast micromethod suggesting it to be the most susceptible species in the current study. However, *P. miliaris* had the highest DNA damage in all enzyme treatments using the modified fast micromethod. This may be explained in that the modified fast micromethod takes a total of 4 h compared with 2.5 h for the unmodified fast micromethod. Previous research showed sea urchins to have robust DNA repair capabilities, with repair initiated and progressing within 1 h [12]. This suggests that the DNA damage detected by the modified fast micromethod was possibly not the maximum DNA damage caused by H_2O_2 , but the net DNA damage after some early DNA repair processes have been initiated. DNA damage susceptibility can be considered as the sum of DNA protection, DNA damage, and biochemical responses to that damage (DNA repair, antioxidant activity), which can vary between cell types (e.g. differences in nuclear or mitochondrial DNA [62]), between species [33], and life strategies (e.g. longevity [63]). Therefore, time differences between the un- and modified fast micromethod could exacerbate the DNA damage response systems (repair and/or antioxidant activity), suggesting that *P. miliaris* may have slightly lower capacity for DNA damage response and higher overall DNA damage susceptibility than the other two species. These interspecific differences in initial vulnerability to DNA damage, and the capacity to initiate a robust and effective DNA repair response, warrants further investigation, and the modified fast micromethod could provide a rapid high-throughput method for multiple species and timepoints.

The current study concludes with further evidence of the potential harm that H_2O_2 can cause in benthic invertebrates in the marine environment around fish farms using H_2O_2 treatment against sea lice. The newly developed FPG and Endo III modified fast micromethod was able to detect and quantify oxidized bases, and differentiate between DNA strand breaks, oxidized purines, and oxidized pyrimidines. Data from the modified fast micromethod indicate that H_2O_2 causes oxidative DNA damage in sea urchin coelomocytes and that it induces predominantly DNA strand breaks, and more oxidized purines than pyrimidines. The three sea urchin species from the same cold-water environment have different levels of H_2O_2 -induced oxidative DNA damage, suggesting differences in DNA damage susceptibilities, or overall DNA damage response systems including capacity for DNA repair. Sea urchin test size, cell concentration, and red cell percentage showed large vari-

ations between individuals and provide comparisons between the temperate species important in coastal ecosystems hosting Atlantic salmon aquaculture industrial activity. Furthermore, the impact of H_2O_2 may be more extensive in other benthic and sessile marine organisms given that sea urchins have a very well-developed immune system with high resilience to DNA damage from environmental stressors [34]. Finally, data from the current study show genotoxic effects on sea urchin coelomocytes from environmentally relevant concentrations of H_2O_2 and indicate H_2O_2 releases from fish farms' potential for genotoxic impact on nontarget marine organisms.

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Data availability

The datasets generated for this study are available on request to the corresponding author.

References

1. Urbina MA, Cumillaf JP, Paschke K et al. Effects of pharmaceuticals used to treat salmon lice on non-target species: evidence from a systematic review. *Sci Total Environ* 2019;649:1124–36.
2. Treasurer JW, Grant A. The efficacy of hydrogen peroxide for the treatment of farmed Atlantic salmon, *Salmo salar* L. infested with sea lice (Copepoda: Caligidae). *Aquaculture* 1997;148:265–75.
3. Bruno DW, Raynard RS. Studies on the use of hydrogen peroxide as a method for the control of sea lice on Atlantic salmon. *Aquac Int J Eur Aquac Soc* 1994;2:10–8.
4. Burrige LE, Lyons MC, Wong DKH et al. The acute lethality of three anti-sea lice formulations: AlphaMax®, Salmosan®, and Interlox®Paramove™50 to lobster and shrimp. *Aquaculture* 2014;420–1:180–6.
5. Ernst W, Doe K, Cook A et al. Dispersion and toxicity to non-target crustaceans of azamethiphos and deltamethrin after sea lice treatments on farmed salmon, *Salmo salar*. *Aquaculture* 2014;424–5:104–12.
6. Refseth GH, Sæther K, Drivdal M et al. Miljørisiko ved bruk av hydrogenperoksid Økotoksikologisk vurdering og grenseverdi for effekt. Akvaplan-niva AS Rapport: 8200. 2016.
7. Giguère PA. Molecular association and structure of hydrogen peroxide. *J Chem Educ* 1983;60:399–401.
8. Frantzen M, Bytingsvik J, Tassara L et al. Effects of the sea lice bath treatment pharmaceuticals hydrogen peroxide, azamethiphos and deltamethrin on egg-carrying shrimp (*Pandalus borealis*). *Mar Environ Res* 2020;159:1–9.
9. Lindahl T. Instability and decay of the primary structure of DNA. *Nature* 1993;362:709–15.
10. Gu X, Sun J, Li S et al. Oxidative stress induces DNA demethylation and histone acetylation in SH-SY5Y cells: potential epigenetic mechanisms in gene transcription in Aβ production. *Neurobiol Aging* 2013;34:1069–79.

11. Kasai H, Nishimura S. Hydroxylation of deoxyguanosine at the C-8 position by ascorbic acid and other reducing agents. *Nucleic Acids Res* 1984;12:2137–45.
12. Reinardy HC, Dharamshi J, Jha AN et al. Changes in expression profiles of genes associated with DNA repair following induction of DNA damage in larval zebrafish *Danio rerio*. *Mutagenesis* 2013;28:601–8.
13. Reinardy HC, Bodnar AG. Profiling DNA damage and repair capacity in sea urchin larvae and coelomocytes exposed to genotoxicants. *Mutagenesis* 2015;30:829–39.
14. Driessens N, Versteijhe S, Ghaddhab C et al. Hydrogen peroxide induces DNA single- and double-strand breaks in thyroid cells and is therefore a potential mutagen for this organ. *Endocr Relat Cancer* 2009;16:845–56.
15. Sancar A, Lindsey-Boltz LA, Ünsal-Kaçmaz K et al. Molecular mechanisms of mammalian DNA repair and the DNA damage checkpoints. *Annu Rev Biochem* 2004;73:39–85.
16. Cao H, Salazar-García L, Gao F et al. Novel approach reveals genomic landscapes of single-strand DNA breaks with nucleotide resolution in human cells. *Nat Commun* 2019;10:1–14.
17. Kuzminov A. Single-strand interruptions in replicating chromosomes cause double-strand breaks. *Proc Natl Acad Sci USA* 2001;98:8241–6.
18. Klungland A, Bjelland S. Oxidative damage to purines in DNA: role of mammalian Ogg1. *DNA Repair* 2007;6:481–8.
19. Evans MD, Dizdaroglu M, Cooke MS. Oxidative DNA damage and disease: induction, repair and significance. *Mutat Res Rev Mutat Res* 2004;567:1–61.
20. Comely CA, Ansell AD. Invertebrate associates of the sea urchin, *Echinus esculentus* L., from the Scottish west coast. *Ophelia* 1987;28:111–37.
21. Kelly MS, Hughes AD, Cook EJ. *Psammechinus miliaris*. In: Lawrence JM (ed.), *Sea Urchins: Biology and Ecology*. UK: Elsevier/London, 2013;38:329–336.
22. Bayed A, Quiniou F, Benrha A et al. The *Paracentrotus lividus* populations from the northern Moroccan Atlantic coast: growth, reproduction and health condition. *J Mar Biol Assoc UK* 2005;85:999–1007.
23. Coteur G, Danis B, Fowler SW et al. Effects of PCBs on reactive oxygen species (ROS) production by the immune cells of *Paracentrotus lividus* (Echinodermata). *Mar Pollut Bull* 2001;42:667–72.
24. Soualili D, Dubois P, Gosselin P et al. Assessment of seawater pollution by heavy metals in the neighbourhood of Algiers: use of the sea urchin, *Paracentrotus lividus*, as a bioindicator. *ICES J Mar Sci* 2008;65:132–9.
25. Smith AB, Kroh A. 2013. Phylogeny of sea urchins. In: Lawrence JM (ed.), *Sea Urchins: Biology and Ecology*. UK: Elsevier/London, 2013:38:1–14.
26. Yorke CE, Page HM, Miller RJ. Sea urchins mediate the availability of kelp detritus to benthic consumers. *Proc R Soc B Biol Sci* 2019;286:1–8.
27. Escobar-Lux RH, Parsons AE, Samuelsen OB et al. Short-term exposure to hydrogen peroxide induces mortality and alters exploratory behaviour of European lobster (*Homarus gammarus*). *Ecotoxicol Environ Saf* 2020;204:1–9.
28. Escobar-Lux RH, Fields DM, Browman HI et al. The effects of hydrogen peroxide on mortality, escape response, and oxygen consumption of *Calanus* spp. *FACETS* 2019;4:626–37.
29. Escobar-Lux RH, Samuelsen OB. The acute and delayed mortality of the northern krill (*Meganyctiphanes norvegica*) when exposed to hydrogen peroxide. *Bull Environ Contam Toxicol* 2020;105:705–10.
30. Campani T, Casini S, Caliani I et al. Ecotoxicological investigation in three model species exposed to elutriates of marine sediments inoculated with bioplastics. *Front Mar Sci* 2020;7:1–9.
31. Schröder HC, Batel R, Schwertner H et al. Fast micromethod DNA single-strand-break assay. *Methods Mol Biol* 2006;314:287–305.
32. Bihari N, Batel R, Jaksic Z et al. Comparison between the comet assay and fast micromethod for measuring DNA damage in HeLa cells. *Croat Chem Acta* 2002;75:793–804.
33. El-Bibany AH, Bodnar AG, Reinardy HC. Comparative DNA damage and repair in echinoderm coelomocytes exposed to genotoxicants. *PLoS One* 2014;9:1–9.
34. Loram J, Raudonis R, Chapman J et al. Sea urchin coelomocytes are resistant to a variety of DNA damaging agents. *Aquat Toxicol* 2012;124–5:133–8.
35. Kushwaha S, Vikram A, Trivedi PP et al. Alkaline, Endo III and FPG modified comet assay as biomarkers for the detection of oxidative DNA damage in rats with experimentally induced diabetes. *Mutat Res Genet Toxicol Environ Mutagen* 2011;726:242–50.
36. Collins AR, Dušinská M, Gedik CM et al. Oxidative damage to DNA: do we have a reliable biomarker? *Environ Health Perspect* 1996;104:465–9.
37. Mustafa SA, Al-Subiai SN, Davies SJ et al. Hypoxia-induced oxidative DNA damage links with higher level biological effects including specific growth rate in common carp, *Cyprinus carpio* L. *Ecotoxicology* 2011;20:1455–66.
38. Schalow BJ, Courcelle CT, Courcelle J. *Escherichia coli* Fpg glycosylase is nonredundant and required for the rapid global repair of oxidized purine and pyrimidine damage in vivo. *J Mol Biol* 2011;410:183–93.
39. Smith LC, Hawley TS, Henson JH et al. Methods for collection, handling, and analysis of sea urchin coelomocytes. *Methods Cell Biol* 2019;150:357–89.
40. Collins AR, Duthie SJ, Dobson VL. Direct enzymic detection of endogenous oxidative base damage in human lymphocyte DNA. *Carcinogenesis* 1993;14:1733–5.
41. Andersson MA, Hellman BE. Different roles of Fpg and Endo III on catechol-induced DNA damage in extended-term cultures of human lymphocytes and L5178Y mouse lymphoma cells. *Toxicol In Vitro* 2005;19:779–86.
42. Pinsino A, Della Torre C, Sammarini V et al. Sea urchin coelomocytes as a novel cellular biosensor of environmental stress: a field study in the Tremiti Island Marine Protected Area, Southern Adriatic Sea, Italy. *Cell Biol Toxicol* 2008;24:541–52.
43. Pinsino A, Matranga V. Sea urchin immune cells as sentinels of environmental stress. *Dev Comp Immunol* 2015;49:198–205.
44. Arizza V, Vazzana M, Schillaci D et al. Gender differences in the immune system activities of sea urchin *Paracentrotus lividus*. *Comp Biochem Physiol A Mol Integr Physiol* 2013;164:447–55.
45. Coates CJ, McCulloch C, Betts J et al. Echinochrome A release by red spherule cells is an iron-withholding strategy of sea urchin innate immunity. *J Innate Immun* 2018;10:119–30.
46. Smith LC, Ghosh J, Buckley KM et al. Echinoderm immunity. *Adv Exp Med Biol* 2010;708:260–301.
47. Service M, Wardlaw A. Echinochrome-A as a bactericidal substance in the coelomic fluid of *Echinus esculentus* (L.). *Comp Biochem Physiol B Comp Biochem* 1984;79:161–5.
48. Matranga V, Pinsino A, Celi M et al. Monitoring chemical and physical stress using sea urchin immune cells. *Echinodermata* 2005;39:85–110.
49. Johnson PT, Chapman FA. Abnormal epithelial growth in sea urchin spines (*Strongylocentrotus franciscanus*). *J Invertebr Pathol* 1970;16:116–22.
50. Branco PC, Borges JCS, Santos MF et al. The impact of rising sea temperature on innate immune parameters in the tropical subtidal sea urchin *Lytechinus variegatus* and the intertidal sea urchin *Echinometra lucunter*. *Mar Environ Res* 2013;92:95–101.
51. Matranga V, Toia G, Bonaventura R et al. Cellular and biochemical responses to environmental and experimentally induced stress in sea urchin coelomocytes. *Cell Stress Chaperones* 2000;5:113–20.
52. Elmhirst R. Habits of *Echinus esculentus*. *Nature* 1922;110:667.
53. Boudouresque CF, Verlaque M. *Paracentrotus lividus*. In: Lawrence JM (ed.), *Sea Urchins: Biology and Ecology*. UK: Elsevier/London, 2013;38:297–327.

54. Schröder HC, Di Bella G, Janipour N et al. DNA damage and developmental defects after exposure to UV and heavy metals in sea urchin cells and embryos compared to other invertebrates. *Prog Mol Subcell Biol* 2005;39:111–37.
55. Le Bouffant R, Cormier P, Cuffe A et al. Sea urchin embryo as a model for analysis of the signaling pathways linking DNA damage checkpoint, DNA repair and apoptosis. *Cell Mol Life Sci* 2007;64:1723–34.
56. Le Bouffant R, Boulben S, Cormier P et al. Inhibition of translation and modification of translation factors during apoptosis induced by the DNA-damaging agent MMS in sea urchin embryos. *Exp Cell Res* 2008;314:961–8.
57. Ghadhab C, Kyrilli A, Driessens N et al. Factors contributing to the resistance of the thyrocyte to hydrogen peroxide. *Mol Cell Endocrinol* 2019;481:62–70.
58. Su Y, Duan J, Ying Z et al. Increased vulnerability of parkin knock down PC12 cells to hydrogen peroxide toxicity: the role of salsolinol and NM-salsolinol. *Neuroscience* 2013;233:72–85.
59. Schlenk D, Rice CD. Effect of zinc and cadmium treatment on hydrogen peroxide-induced mortality and expression of glutathione and metallothionein in a teleost hepatoma cell line. *Aquat Toxicol* 1998;43:121–9.
60. Zhu C-Y, Loft S. Effects of brussels sprouts extracts on hydrogen peroxide-induced DNA strand breaks in human lymphocytes. *Food Chem Toxicol* 2001;39:1191–7.
61. Gabbianelli R, Falcioni ML, Cantalamessa F et al. Permethrin induces lymphocyte DNA lesions at both Endo III and Fpg sites and changes in monocyte respiratory burst in rats. *J Appl Toxicol* 2009;29:317–22.
62. Sawyer DE, Roman SD, Aitken RJ. Relative susceptibilities of mitochondrial and nuclear DNA to damage induced by hydrogen peroxide in two mouse germ cell lines. *Redox Rep* 2001;6:182–4.
63. Ungvari Z, Ridgway I, Philipp EER et al. Extreme longevity is associated with increased resistance to oxidative stress in *Arctica islandica*, the longest-living non-colonial animal. *J Gerontol Ser A Biol Sci Med Sci* 2011;66A:741–50.